

1 Temporal dynamics of microbial transcription in wetted 2 hyperarid desert soils

3 Carlos León-Sobrino^{1,2}, Jean-Baptiste Ramond^{1,3}, Clément Coclet¹, Ritha-Meriam
4 Kapitango⁴, Gillian Maggs-Kölling⁴, and Don A. Cowan^{1,*}

5 ¹ Centre for Microbial Ecology and Genomics, Department of Biochemistry,
6 Genetics and Microbiology, University of Pretoria, Pretoria, South Africa.

7 ² Current address: The Novo Nordisk Foundation Center for Biosustainability,
8 Technical University of Denmark, 2800 Kgs Lyngby, Denmark.

9 ³ Departamento Genética Molecular y Microbiología, Pontificia Universidad
10 Católica de Chile, Chile.

11 ⁴ Gobabeb - Namib Research Institute, Walvis Bay, Namibia.

12 * To whom correspondence should be addressed: don.cowan@up.ac.za

13 **Abstract**

14 Rainfall is rare in hyperarid deserts but, when it occurs, it triggers large biological
15 responses which are considered to be essential for the long-term maintenance of
16 biodiversity. In such environments, microbial communities have major roles in nutrient
17 cycling, but their functional responses to short-lived resource opportunities are poorly
18 understood. We used whole community metatranscriptomic data to demonstrate
19 structured and sequential functional responses in desiccated desert soils to a
20 simulated rainfall event over a seven-day time frame. Rapid transcriptional activation

21 of Actinobacteria, Alpha-proteobacteria and phage transcripts was followed by a
22 marked increase in protist and myxobacterial activity, before returning to the original
23 state. In functional terms, motility systems were activated in the early phases, whereas
24 competition-toxicity systems increased in parallel to the activity from predators and
25 the drying of soils. The dispersal-predation dynamic was identified as a central driver
26 of microbial community responses to watering. Carbon fixation mechanisms that were
27 active under dry condition were rapidly down-regulated in wetted soils, and only
28 reactivated on a return to a near-dry state, suggesting a reciprocal balance between
29 carbon fixation and fixed-carbon acquisition processes. Water addition induced a
30 general reduction in the transcription of stress response genes, most prominently
31 HSP20, indicating that this chaperone is particularly important for life in desiccated
32 ecosystems. Based on these data, we propose a rainfall response model for desert soil
33 microbiomes.

34 ***Introduction***

35 Arid lands cover approximately one third of the terrestrial surface (Laity, 2009).
36 Aridity, as a result of very low precipitation and high potential evapotranspiration
37 rates, severely limits biomass production and promotes oligotrophy (Delgado-
38 Baquerizo et al., 2013; Maestre et al., 2015), eventually leading to habitat
39 fragmentation where productivity is concentrated in sheltered "islands" (Pointing and
40 Belnap, 2012; Schlesinger et al., 1995). Microorganisms in open soils outside of these
41 privileged micro-environments are subjected to an intense desiccation stress which
42 limits their biological activity (Lebre et al., 2017). However, desert edaphic
43 microorganisms are regarded as important for the long-term fertility of soils and for

44 post-rainfall grass germination (Delgado-Baquerizo et al., 2016). The Namib Desert, on
45 the south-western coast of Africa, is the oldest continuously hyperarid desert in the
46 world and provides an excellent model of a stable hyperarid ecosystem (Seely et al.,
47 2008). It is generally accepted that much of the microbiome in such environments
48 exists in a state of dormancy (Bär et al., 2002; Lebre et al., 2017), but responds with
49 dramatic changes in structure and activity when water becomes available (Armstrong
50 et al., 2016; Austin et al., 2004; Collins et al., 2014; Garcia-Pichel and Pringault, 2001;
51 Štoviček et al., 2017). Recent studies have, however, shown that some transcriptional
52 activity is detectable in hyperarid soil microbial communities during prolonged dry
53 periods (Gunnigle et al., 2017; León-Sobrino et al., 2019; Schulze-Makuch et al., 2018).
54 Since arid lands represent a large part of the terrestrial surface of the Earth and are
55 currently expanding through desertification processes (Huang et al., 2016), a deeper
56 understanding of these pulsed phenomena is of global relevance.

57 Research on desert soil microbial ecology has primarily focused on bacterial
58 communities, suggesting that extreme abiotic pressures, such as high temperature,
59 desiccation stress and UV radiation are dominant drivers of both the diversity and
60 function of bacterial communities (Frossard et al., 2015; Johnson et al., 2017; León-
61 Sobrino et al., 2019; Mandakovic et al., 2018; McHugh et al., 2017; Ronca et al., 2015;
62 Scola et al., 2018; Vikram et al., 2016). By comparison, virus/phage communities in hot
63 desert soils have been relatively under-studied (Trubl et al., 2020; Williamson et al.,
64 2017), despite the fact that they represent one of the most abundant and genetically
65 diverse entities on Earth and are inferred to play central ecological roles in
66 biogeochemical nutrient cycling in diverse environments (Adriaenssens et al., 2017;

67 Bezuidt et al., 2020; Luo et al., 2020; Zablocki et al., 2018). Furthermore, phages (i.e.,
68 viruses that infect bacteria) have been shown to affect community turnover and
69 resource availability in soils and to globally drive microbial evolution via horizontal
70 gene transfer (Rodriguez-Valera et al., 2009; Trubl et al., 2019; Williamson et al., 2017).

71 In this study, a gravel plain soil plot in the central Namib Desert was subject to an
72 artificial rainfall pulse of 30 L/m², the approximate average annual rain received in this
73 region (25 mm) (Eckardt et al., 2013) and sufficient to stimulate plant germination
74 (Seely et al., 2008). The short-term responses of the soil microbiome were monitored
75 using whole-transcriptome analysis of gene function. Based on this analysis, we
76 propose a structured water response model with differentiated phases and trophic
77 interactions.

78 ***Materials and methods***

79 Surface soils were collected from two adjacent 3.5 x 3.5 m plots (approx. 10m
80 separation) in the central Namib Desert gravel plains (23° 33' 18" S, 15° 3' 20" E). The
81 control plot remained dry, while the experimental plot was manually watered (30L/m²,
82 simulating a 30 mm rainfall event) at T₀ (26th April 2017 10:00 AM WAT/UTC+1), using
83 a synthetic "Namib Desert rain" solution, prepared from ultrapure DNA-free water
84 supplemented with a defined salt mixture (Frossard et al., 2015). Plots were
85 subdivided into 0.5 x 0.5 m quadrats and randomly sampled, in triplicate, at specified
86 times after water addition (Supplementary Figure S1). 20 g samples of surface (0-5 cm)
87 soil were collected at 10 minutes; 1, 3 and 7 hours; and 1, 3 and 7 days after simulated
88 rainfall, preserved immediately in RNAlater solution (Sigma-Aldrich, St. Louis MO, USA)
89 and subsequently frozen at -20°C prior to RNA extraction. Additional 200 g soil samples

90 were collected from the same locations into WhirlPak™ bags (Nasco, Fort Atkinson WI,
91 USA) and frozen for subsequent soil chemistry analysis.

92 The water content of soil samples (sieved to <2 mm particles) was measured
93 gravimetrically for three days after watering. Soil silt, sand, and clay compositions
94 were measured by the hydrometer method (Gee and Bauder, 1986). Particle size
95 distributions were determined by sieve separation (> 1000 µm, > 500, > 250, > 100, >
96 53 and < 53 µm). pH, electrical conductivity, Na, Cl, K, Ca, Mg, NO₃, NH₄ and P
97 composition were analysed at Bemlab (Pty) Ltd. (Strand, Western Cape, South Africa)
98 using standard protocols. Soil organic carbon percentage was measured using the
99 Walkley-Black test (Walkley, 1935).

100 RNA was extracted following the protocols described in León-Sobrinó *et al.* 2019
101 (León-Sobrinó *et al.*, 2019) from triplicate soil samples. In order to mitigate the effect
102 of chemical variation, the two biological replicates most similar to the average
103 chemical composition of all sampled soils were selected for RNA extraction. Stranded,
104 rRNA-depleted libraries were prepared with the ScriptSeq Complete Gold Kit
105 (Epidemiology) (Illumina, San Diego, USA) and 150 bp paired-end sequences were read
106 on a HiSeq4000 platform (Illumina).

107 Sequencing outputs were processed using the BBtools suite v. 38.26 (Bushnell *et*
108 *al.*, 2017) (<https://sourceforge.net/projects/bbmap/>). Read ends below a quality Phred
109 value of 20 were trimmed using *BBDuk*; rRNA and human RNA sequences were
110 identified and removed using SILVA v. 111 (July 2012) (Quast *et al.*, 2012) and 5SRNA
111 (Szymanski *et al.*, 2016) databases and a curated human genome reference assembly
112 hg19 (<https://drive.google.com/file/d/0B3lIHR93L14wd0pSSnFULUhcUk>) (bbmap,

113 2018) following recommended protocols. Optical duplicates generated by the
114 patterned sequencing flowcell were removed using the *Clumpify* function from the
115 BBtools suite, setting the distance cut-off to 2,500 pixels. Transcript assembly was
116 performed using transAbyss v.2.0.1 (Robertson et al., 2010) for each library. Individual
117 assemblies were merged with the same software (*transabyss-merge* function) to
118 generate a reference metatranscriptome.

119 Contigs were annotated at the Integrated Microbial Genomes and Microbiomes
120 (IMG/M) server (Huntemann et al., 2015) (<https://img.jgi.doe.gov/>). Predicted protein
121 products from genes were also analysed against the Conserved Domain Database
122 (CDD) using Delta Blast (e-value threshold 10^{-4}) (Boratyn et al., 2012; Marchler-Bauer
123 et al., 2015). Contig taxonomy was determined by a consensus among all individually
124 classified genes, requiring a *quorum* of >50% at each taxonomic level.

125 Reads were aligned to the reference assembly using BBmap (Bushnell et al., 2017)
126 and reads for annotated regions in each library were counted using FeatureCounts v.
127 1.6.3 (Liao et al., 2014). The assembled counts matrix was aggregated along functional
128 and/or taxonomic categories as required for each analysis.

129 Differential transcription along the time series was analysed with the DESeq2 v.
130 1.14 package in R (Love et al., 2014), comparing data from the control (dry) soil
131 samples with those from the experimental (wetted) samples from the same point in
132 the time-series. Transcripts that significantly diverged in abundance from the control
133 at any given time were considered up-regulated in response to watering (adjusted p-
134 value ≤ 0.05 for the likelihood ratio test). Transcripts per million (TPM) of ribosomal
135 protein genes as a fraction of the total for their respective taxonomic group (Rp:T)

136 were employed to estimate absolute activity of each group along the time course,
137 following the premise that ribosome densities in a cell relate to metabolic activity and
138 growth rates (Bosdriesz et al., 2015; Bremer and Dennis, 1996).

139 Viral contigs were identified from assembled transcriptomes using VirSorter v. 1.0.3
140 (Roux et al., 2015) and the virome database on the iVirus platform hosted by CyVerse
141 (Bolduc et al., 2017). Only contigs >1 kb, and classified as categories 1, 2, 4, and 5 were
142 considered (phages and prophages identified with the “pretty sure” and “quite sure”
143 qualification). To calculate the relative abundances of the different viral contigs in each
144 transcriptome, quality filtered metatranscriptomic reads were mapped back to the
145 viral contigs with Bowtie2 v. 2.2.6, using the default parameters (Langmead and
146 Salzberg, 2012). The output SAM files were converted into BAM files, sorted and
147 indexed, using SAMtools (Li et al., 2009). A python script was used to generate a
148 coverage profile of viral contig abundances across samples.

149 ORFs were predicted within putative viral contigs using Prodigal (Hyatt et al.,
150 2010). Transcripts per million (TPM) putative viral protein genes were employed to
151 normalize the final ‘viral OTU’ (vOTU) values for each sample. Predicted protein
152 sequences were clustered with proteins from viruses in the NCBI ViralRefSeq-
153 prokaryotes-v85 based on all- versus-all BLASTp search with an E value of 1×10^{-3} , and
154 clusters were defined with the Markov clustering algorithm and processed using
155 vConTACT2 (Bin Jang et al., 2019). The stringency of the similarity score was evaluated
156 through 1,000 randomizations by permuting protein clusters or singletons (proteins
157 without significant shared similarity to other protein sequences) within pairs of
158 sequences having a significance score ≤ 1 (negative control). Subsequently, pairs of

159 sequences with a similarity score > 1 were clustered into viral clusters with the Markov
160 clustering algorithm using an inflation value of 2. The resulting gene-sharing network
161 from vConTACT2 classification was visualized with Cytoscape software v. 3.7.0 (Smoot
162 et al., 2011). Reference sequences from RefSeq database that co-clustered with the
163 putative viral sequences were used to predict viral taxonomy.

164 **Results**

165 ***Site characteristics and taxonomic analysis***

166 The sample site (Figure 1A and 1B) was a characteristic Namib Desert calcrete
167 gravel plain (Gombeer et al., 2015) with high sand composition (92 ± 1.6 %) and very
168 low organic carbon content (0.04%) (Supplementary Table S1). Soil chemical
169 composition was relatively homogeneous in all sampled sectors and between the
170 sample and control sites (Supplementary Table S1). Gravimetric water content
171 measurements showed that more than half of the water content in the surface soil (0-
172 5 cm) was lost 24 hours after the simulated 30 mm rainfall. After 3 days, the water
173 content was similar to that of the dry control (Figure 1C).

174 Transcript read assembly yielded a consensus metatranscriptome of 208.95 Mb in
175 which 378 802 coding regions were annotated, including 372 044 predicted protein-
176 coding genes. On average, for the 24 sequenced libraries, 61.7% of reads could be
177 aligned back to contigs. Functions were predicted for 29.9% of the protein-coding
178 genes using the KEGG database (Kanehisa et al., 2016), and 51.6% using the CDD.
179 56.8% of contigs were taxonomically classified at phylum level, and 56.3% down to
180 family level.

181 A functional and taxonomic analysis of the transcriptional profiles revealed large
182 differences in gene expression levels between treatment and control soil
183 transcriptomes within 10 minutes after watering (Figure 2, Supplementary Figure S2).
184 Communities from dry soils were characterized by stable (i.e., largely unchanged)
185 transcription over the 7-day experimental period. In contrast, microbial communities
186 in the watered soil plot underwent an immediate and large change in gene expression
187 (Figure 2A) that progressively returned to basal expression levels within 7 days from
188 the wetting event. Taxonomy-only transcriptional profiles followed this cyclical pattern
189 (Figure 2B) but were less robust than KO gene categories in representing the changes
190 in the soil microbiome induced by soil wetting.

191 In both watered and dry soils, Actinobacteria and Proteobacteria were the most
192 abundant phyla (average 40% and 16% in all samples, respectively, see Supplementary
193 Table S2). Strikingly, transcripts from Nitrososphaeria, a class of ammonia oxidizing
194 Thaumarchaeota, were the third most abundant (average 14% of classified TPM in all
195 samples), suggesting that these taxa play an important role in the microbial
196 community and in local nitrogen cycling. Protist transcripts (Oligohymenophorea class
197 and Dictyostellales), which were rare in dry soil samples (<0.5%), increased to >4%
198 within 7 hours of soil watering. Conversely, transcripts of fungal origin were more
199 common in dry soils (1% of classified TPM) than in watered ones.

200 The production of ribosomes in each taxonomic group was analysed, as an
201 indicator of their cellular metabolic activity, by quantifying ribosomal protein
202 transcripts as a fraction of the total (Rp:T) (Bosdriesz et al., 2015; Bremer and Dennis,
203 1996). Ribosomal protein transcripts constituted a stable proportion for each taxon in

204 dry control samples, typically <4% TPM (Figure 3). After water addition, however, Rp:T
205 increased for all major taxa, peaking at characteristic times over the course of the
206 experiment and returning to basal values by the end of the 7 day period.
207 Actinobacteria, Alpha-proteobacteria and Chloroflexi were examples of ‘early-
208 activation’ groups, with Rp:T values peaking within the first hour after water addition.
209 Gemmatimonadetes experienced the largest relative increase in Rp:T, despite being a
210 minor component of the soil microbiome (<1% TPM). A delayed rise in activity, peaking
211 at 7 hours after the simulated rainfall, was evident for Delta-proteobacteria, protists
212 and Bacteroidetes, especially those belonging to the Cytophagia class, which
213 constituted the principal ‘late-activation’ group. The Pezizomycetes, the most active
214 fungal class, also showed a slower response to water addition. Intermediate response
215 patterns, with maximal Rp:T values at 3 hours after water addition, were observed for
216 taxa such as Planctomycetes. Thaumarchaeal Rp:T ratios showed only moderate
217 changes after watering and a relatively even distribution throughout the experiment.

218 ***Temporal changes in core cellular functions***

219 Watering induced a general reduction of stress resistance gene transcripts in the
220 microbial community immediately after watering (e.g., *groEL*, *dnaK*, *clp*, *tre*). Trehalose
221 synthesis genes (*tre*), which drive solute accumulation under conditions of water stress
222 (Lebre et al., 2017), declined in relative transcription in watered soils. Transcripts for
223 the major chaperones *groEL* and *dnaK*, and the *clp* protease involved in protein
224 misfolding control were also reduced. The most conspicuous change in core stress
225 resistance systems was, however, that of the heat-shock protein HSP20 genes
226 (KO13993, CD223149 or CD278440). Genes for this ATP-independent chaperone

227 experienced the largest and most consistent transcript reduction across all significant
228 (average ≥ 1 % of transcripts) bacterial taxa after watering (Figure 4), suggesting that
229 HSP20 represents an archetypal "desiccation adaptation" gene.

230 Motility related transcripts were rapidly affected by soil wetting (from 10 minutes
231 after water addition). Flagellar genes were up-regulated within the first hour in the
232 dominant Actinobacteria and Alpha-proteobacteria (Figure 4) and in less dominant
233 taxa such as Planctomycetes and Firmicutes. Type IV pilus genes, involved in gliding
234 motility, were also rapidly upregulated after water addition, most notably in
235 Actinobacteria. These responses suggest that rapid dispersal is a key behavioural
236 response to soil wetting, particularly in Actinobacteria.

237 A significant up-regulation of Eukaryotic adhesion and cytoskeletal components,
238 myosin, actin and tubulin, was evident 3 to 7 hours after soil wetting (Figure 4), but
239 was limited to non-fungal taxa, particularly in Oligohymenophorea (ciliate) and
240 Dictyosteliales (slime mould).

241 Another highly significant increase in transcriptional activity observed several
242 hours after water addition involved interbacterial competition and predation genes.
243 These included type VI secretion systems (T6SS) and vibriolysin encoding genes from
244 the order Myxococcales (Delta-proteobacteria), and T6SS and serralysin encoding
245 genes from Alpha-proteobacteria. Heightened T6SS transcription was also observed in
246 several other bacterial taxa, such as Planctomycetes, Gemmatimonadetes and
247 Gamma-proteobacteria, suggesting that antagonistic interactions are of significant
248 ecological importance in newly wetted soils.

249 ***Biogeochemical cycles***

250 *Carbon utilization:* Despite the observation that microbial photosynthetic processes
251 are water-limited in hyperarid soils (Tracy et al., 2010; Warren-Rhodes et al., 2006),
252 transcript data did not show a significant increase in cyanobacterial gene expression
253 after water addition, and both overall cyanobacterial transcript abundance and Rp:T
254 ratios remained low (Figure 3, Supplementary Table S2). Furthermore, core carbon
255 fixation genes, including RuBisCO and carboxylases genes from chemoautotrophic
256 pathways, consistently showed reduced relative transcription in wetted soils.
257 Thaumarchaeal carbon fixation was seemingly also affected by watering, reflected in
258 the transient inhibition of 4-hydroxybutyryl-CoA dehydratase from the 3-
259 hydroxypropionate/4-hydroxybutyrate cycle following inundation. These data are
260 indicative of an inhibition of both photosynthetic and non-photosynthetic carbon
261 fixation mechanisms immediately after soil saturation (Fig S3A). However, expression
262 of carbon fixation genes increased by the end of the 7-day experimental period, when
263 soil water content had returned to basal levels (Figure 4, Supplementary Figure S3A).

264 Dramatic increases in CO₂ emissions from newly wetted desert soils, attributed to
265 degradation of dissolved organic matter, have been widely reported (Austin et al.,
266 2004). Soil chemistry, however, showed no significant reduction in soil Total Organic
267 Carbon (Supplementary Table S1), possibly due to limited sensitivity of the employed
268 method at the very low values present in these soils. In our transcript data, indicators
269 of biomass degradation, including carbohydrate and peptide transport systems (e.g.,
270 *rbsB*, *xyIF*, and *livK*, *dppA*) significantly increased for most taxa immediately after water
271 addition. The Bacteroidetes phylum, dominated by the Cytophagales, showed dramatic

272 up-regulation of subtilisin protease transcripts and components of the protein and
273 carbohydrate import machinery (*ragA/susC* CD274948; *susD* CD185760), suggesting an
274 important role of this group in the turnover of organic matter in the ecosystem.

275 *Nitrogen and phosphate utilization:* Transcript data suggested that Thaumarchaea
276 were the main drivers of nitrogen cycling in dry soils. Ammonia monooxygenase and
277 NO-forming nitrite reductase (*amo* and *nirK*) transcripts, originating largely from
278 Thaumarchaea, were among the most abundant overall. Transcripts associated with
279 both processes declined after soil watering (Figure 5A, Supplementary Figure S3B). All
280 other detected transcripts implicated in ammonification, such as nitrate and nitrite
281 reductases *narGH* and *nirAD* (mainly transcribed by Nitrospirae and Actinobacteria,
282 respectively), and nitric oxide reductase (*norC*), were also reduced after watering.
283 Conversely, peptide transporter transcripts (*livKH*, *dppA*) significantly increased in
284 response to water addition for several taxa (Actinobacteria, Alpha-, Beta- and Delta-
285 proteobacteria), suggesting that organic nitrogen dominates N acquisition processes in
286 newly wetted soils.

287 Changes in phosphorus acquisition pathways were also identified. Transcript data
288 suggested that phosphate acquisition after water addition was dominated by the
289 proteobacterial *ugpB* sn-glycerol 3-phosphate (G3P) transport system subunit,
290 particularly in Alpha-proteobacteria, followed by the *pst* inorganic phosphate
291 transporter (Figure 5B). By contrast, phosphonate transport (*phn*) gene transcription
292 was reduced after soil wetting.

293 ***Analysis of viral transcripts***

294 A total of 68 contigs were characterized as being of viral origin using VirSorter
295 (Supplementary Figure S4). In both dry and watered soil samples, viral contigs were
296 significantly lower than bacterial contigs (ANOVA, $p < 0.01$), representing an average of
297 $0.12 \pm 0.07\%$ TPM of viral protein transcripts. Over the period of the experiment, viral
298 reads exhibited a significant temporal pattern, where read numbers remained low in
299 all dry soil samples but exhibited a bimodal response in wetted soils (Figure 6). An
300 initial very rapid increase in viral RNAs (approx. 2.2-fold at the first post-watering
301 sample point: 10 min) was followed by a 12-fold decline over 7 hours. A second
302 increase of viral read numbers (approx. 6-fold) occurred between 7 hours and 7 days.

303 To investigate the diversity of the ‘active’ viral population, we used a genome-based
304 network analysis of the shared protein content with the prokaryotic viral genomes
305 (RefSeq v85). This analysis grouped 35 viral contigs into viral clusters (VCs) while 33
306 viral contigs remained non-clustered and were considered as singletons (Figure 7). In
307 the network, 10 VCs containing viral contigs from our study were predicted, 7 of which
308 did not belong to VCs with RefSeq virus genomes but instead clustered together into
309 novel VCs, and 3 of which could be assigned taxonomy at the family level (Figure 7) as
310 members of the *Caudovirales* (*Siphoviridae* and *Leviviridae*).

311 **Discussion**

312 Studies of the microbial ecology of desert edaphic niches have tended to focus on
313 biological ‘hotspots’: hypoliths, soil crusts, and soils in the immediate vicinity of plants
314 (e.g., Pointing and Belnap, 2012; Ramond et al., 2008; Marasco et al., 2018). While the
315 use of high-throughput sequencing techniques (Crits-Christoph et al., 2013; Fierer et
316 al., 2012, 2007; León-Sobrino et al., 2019; Jordaan et al., 2020; Vikram et al., 2016) has

317 greatly expanded of knowledge of the microbial ecology of these niches, most studies
318 have centred on microbial diversity rather than functionality. Total RNA sequencing is
319 therefore a valuable tool to monitor microbial functionality at a high temporal
320 resolution, particularly since mRNA is only generated by active organisms and is
321 ephemeral, leaving little to no legacy signal (León-Sobrino et al., 2019; Rajeev et al.,
322 2013; Steven et al., 2018).

323 Previous studies have clearly demonstrated a long-term compositional and
324 functional adaptation of Namib Desert edaphic and hypolithic microbial communities
325 to abiotic factors, particularly water input regimes histories (e.g., fog vs rain inputs;
326 Cowan et al, 2020; Frossard et al., 2015; Ramond et al., 2018; Scola et al., 2018;
327 Stomeo et al., 2013;). The location for this study site was situated in the central hyper-
328 arid zone of the Namib Desert, which receives little input water from either rain or fog
329 (Eckardt et al., 2013). The location should therefore optimise the responsiveness of the
330 soil microbiome to water input (Seely et al., 2008).

331 **The HSP20 chaperone is important for microbial life in desiccated soils**

332 One of the projected effects of water addition was an apparent reduction in
333 cellular stress, suggested by the immediate down-regulation in transcription of stress-
334 resistance genes. The most conspicuous of these changes was the abrupt reduction in
335 transcription - across all major bacterial taxa - of the small ATP-independent heat-
336 shock protein HSP20. This chaperone has been characterized as a broad-spectrum
337 bacterial stress resistance mechanism (Bepperling et al., 2012; Haslbeck and Vierling,
338 2015).

339 **Desert soil microbial community members are sequentially activated after a water**
340 **event**

341 Data from the control (unwatered) site confirmed the presence of a diverse and
342 functionally active microbial community in desiccated desert soils (see also, Gunnigle
343 et al., 2017; Jordaan et al., 2020; León-Sobrino et al., 2019; Schulze-Makuch et al.,
344 2018). However, our data suggest a level of remarkable "metabolic readiness", with a
345 dramatic increase in transcription associated with previously inactive taxa occurred
346 within 10 minutes after water addition. We note that transcriptional response rates
347 may be considerably faster, given that 10 minutes was the first sampling time-point. In
348 polyextreme hyperarid desert soils, where most (micro)organisms remain in a state of
349 metabolic dormancy, such a rapid and opportunistic response to the sudden
350 availability of water is clearly an adaptive advantage to access and utilize more
351 favourable, and newly available, ecological substrates and niches.

352 Water addition led to a general increase in the relative abundance of ribosomal
353 protein transcripts (Rp:T), interpreted as an increase in cellular activity (Bosdriesz et
354 al., 2015; Bremer and Dennis, 1996). Cellular activity levels returned to basal (control)
355 levels within 7 days, in parallel with the desiccation of the soil samples. These data are
356 consistent with the paradigm that desert ecosystems and their indigenous microbiota
357 are both resilient and water-pulse responsive (e.g., Belnap et al., 2005; Noy-Meir,
358 1973; Armstrong et al., 2016).

359 Interestingly, various groups of taxa reached maximum Rp:T values at different
360 times, suggesting a controlled pattern of functionality reminiscent of a stepwise model
361 where ecosystem functions gradually evolve as a function of the duration and intensity

362 of the water pulses (Schwinning and Sala, 2004). The most immediate microbial
363 response was characterized by transcription of genes implicated in the motility
364 apparatus (type IV pili in Alpha-proteobacteria and flagella in Actinobacteria). It has
365 been previously noted that one of the main impacts of water inundation is increased
366 soil particle connectivity, providing access to new niches and solubilized nutrients
367 (Schimel, 2018). Actinobacteria and Alpha-proteobacteria have been reported as the
368 dominant active taxa in desiccated soils (León-Sobrinó et al., 2019), possibly uniquely
369 positioned to access new and more favourable niches during periods of inter-
370 connection associated with the water-saturated state.

371 A significant, but delayed, transcriptional activation was observed in the non-fungal
372 microbial eukaryotes (e.g., protists) and Deltaproteobacteria; i.e., 3 to 7 hours after
373 water addition. The former were mostly characterized by structural gene transcripts
374 from the cytoskeleton, a generic indication of overall cellular activity (cell motility
375 and/or cell division). Up-regulated Deltaproteobacterial transcripts were
376 predominantly derived from the Myxococcales, an order of well-known predatory
377 bacteria (Jurkevitch and Davidov, 2007; Shimkets et al., 2006). The dramatic increase in
378 protist and myxobacterial activity is strongly suggestive of predatory behaviour (Thiery
379 and Kaimer, 2020), possibly triggered by increases in prey abundance (i.e.,
380 Actinobacteria and Alphaproteobacteria populations) rather than just by soil
381 rehydration.

382 **Namib Desert fungal and cyanobacterial soil populations are not activated by water**

383 Among the edaphic taxa that were essentially non-responsive to water addition,
384 we particularly identified the Cyanobacteria and most Fungi, with almost none of their
385 genes being differentially up-regulated at a statistically significant level.

386 We anticipated that water addition would trigger a significant and rapid increase in
387 primary production markers linked to cyanobacterial and photosynthetic activity, as
388 previously observed in biological soil crusts and hypolithic communities (Angel and
389 Conrad, 2013; Pringault and Garcia-Pichel, 2004; Rajeev et al., 2013; Steven et al.,
390 2018; Warren-Rhodes et al., 2006). The almost complete absence of water input-
391 related activation of cyanobacterial functionality suggests that primary productivity in
392 hyperarid soils may not be driven by cyanobacteria and is consistent with previous
393 observations showing that hypolithons (and maybe other cryptic communities) are the
394 foundation of productivity after rain events in the Namib Desert (Ramond et al., 2018).

395 **Desert edaphic viral communities change over time after a wetting event**

396 The extent to which microbial communities in desert soils are influenced by phage
397 remains largely unexplored (Fancello et al., 2013; Zablocki et al., 2016; 2017). This
398 study provides one of the first temporal assessments of active phage in a desert
399 edaphic environment (Zablocki et al., 2016). In parallel with the rapid changes in
400 bacterial metatranscriptomic patterns, the phage population responded within 10 min
401 after water addition, followed by a sharp decrease and a secondary increase broadly
402 correlating with the 'late response' fraction of the soil bacterial community.

403 We propose two possible mechanisms which could lead to the bimodal response of
404 viral communities in wetted soils: a direct response to increased predation and/or a

405 *kill-the-winner* model as observed in aquatic ecosystems (e.g., Shapiro et al., 2010).
406 Microbial taxa in a 'biologically hostile' environment may be forced to allocate
407 resources to energetically expensive defence mechanisms, potentially resulting in
408 lowered defence against phage and elevated susceptibility to infection (Chen and
409 Williams, 2012; Friman and Buckling, 2013; Örmälä-Odegrip et al., 2015). The bacterial
410 host's capacity to maintain the 'arms race' dynamic (i.e., co-escalation of host
411 resistance and parasite infectivity) may also be degraded in the presence of predators
412 (Friman and Buckling, 2013).

413 We note that most of the viral sequences identified in this study were unclassified
414 suggesting, as previously observed (Scola et al., 2018), that hot desert soils may
415 harbour a substantial and potentially novel phage genomic and taxonomic diversity.

416 **Water pulses shift microbial C, N and P nutrient utilization patterns**

417 It was anticipated that water addition would trigger a significant increase in
418 primary productivity, since photosynthetic processes are highly sensitive to water
419 activity (Brock, 1975; Steven et al., 2018; Warren-Rhodes et al., 2006). However,
420 markers for photosynthetic and chemoautotrophic carbon fixation (the latter being
421 active in desiccated periods (León-Sobrino et al., 2019; Sghaier et al., 2016)), were
422 significantly suppressed. This suggests that open soil communities may be dependent
423 on carbon input from alternative sources, such as sporadic vegetation growth and/or
424 productive cryptic niches such as hypolithons and biological soil crusts (Armstrong et
425 al., 2016; Ramond et al., 2018). Nitrogen cycling genes, particularly those involved in
426 inorganic nitrogen acquisition (i.e., nitrate, via nitrate reductases), were also
427 downregulated after water addition. This was also surprising as active N mineralization

428 and N loss processes are often increased in arid soils in response to rainfall (Austin et
429 al., 2004), and desert soils are particularly rich in nitrate (Graham et al., 2008;
430 Walwood et al., 2003). However, a metabolic switch to nitrogen acquisition from
431 organic substrates was strongly suggested by the upregulation of peptide transporter
432 genes, mirroring the situation observed for carbon acquisition. The transient reduction
433 in autotrophic C and N fixation after watering may be explained in terms of energy
434 efficiency, where the sudden availability of 'energy-rich' substrates provides a
435 favoured heterotrophic resource over energetically expensive autotrophic processes
436 (Fuchs, 2011).

437 The addition of water triggered an upregulation of genes involved in inorganic
438 phosphate transport and a downregulation of those implicated in organic phosphonate
439 acquisition. We speculate that the solubilization of inorganic phosphate from soil
440 particles displaces phosphonates as the preferred P source (Schowanek and
441 Verstraete, 1990). Noticeable exceptions were the Alpha-proteobacterial taxa, which
442 apparently favour organic G3P as a preferred P source, both in desiccated soils and
443 after wetting (León-Sobrino et al., 2019).

444 **A conceptual response model of desert soil edaphic microbial communities to water**

445 From a composite analysis of our metatranscriptomic data, we propose a rainfall
446 response model for desert soil microbiomes (Figure 8). Immediately after soil wetting
447 (≤ 10 min), some bacterial taxa (particularly Actinobacteria and Alpha-proteobacteria,
448 that show significant chemoautotrophic capacity in dry soils (León-Sobrino et al.,
449 2019)) reduce autotrophic carbon fixation activities, activate cellular uptake
450 mechanisms and engage in dispersal using both natatory (flagella) and gliding (type IV

451 pili) mechanisms, presumably in order to colonize new niches and access new
452 substrate resource pools. There is a concomitant increase in active phage particles.

453 Following this rapid dispersal burst, approximately 3 to 7 hours after water
454 addition predatory and saprophytic microbial taxa are activated. These predators
455 include eukaryotes, especially ciliates (Oligohymenophorea class), Dictyostellid
456 amoebae, Delta-proteobacteria (myxobacteria) and Bacteroidetes (Cytophagia class).
457 Simultaneously, and presumably in response to the activation of predators, several
458 bacterial groups upregulate the transcription of defensive systems, most notably T6SS.
459 A second wave of viral infections appears to follow the emergence of eukaryotic
460 predators.

461 In the final stage of the wetting-drying cycle (approx. 7 days after water addition),
462 when soils are effectively dehydrated to pre-watering levels, autotrophic carbon
463 fixation processes and nitrogen cycling are reactivated in the bacterial community,
464 along with certain desiccation stress resistance mechanisms, most especially HSP20.

465 **Conclusion**

466 In this 7 day *in situ* metatranscriptomics study, we have observed the dominant
467 microbial processes induced by 'rainfall' on hyperarid soils and successfully followed a
468 complete cycle of microbial activation and subsequent inactivation as soils and their
469 microbial communities returned to their basal desiccated state and transcriptional
470 status. We show strong evidence of short-term temporal succession and, by
471 implication, tightly regulated processes in desert edaphic communities after water
472 addition, which favour heterotrophy over autotrophy. Furthermore, we suggest that

473 dispersal-predation cycles, in which both phage and predators play active roles, are
474 important driving forces in shaping community composition within a wetting-drying
475 regime in hyper-arid desert soils.

476 ***Declarations***

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484 ***Data availability***

485 The dataset supporting the conclusions of this article is available in the
486 ArrayExpress repository, <https://www.ebi.ac.uk/arrayexpress/> (accession no. E-MTAB-
487 9439). Reference metatranscriptome assembly and annotations can be accessed at the
488 IMG/M repository, <https://img.jgi.doe.gov/> (GOLD Analysis Project Id Ga0326365)

489 ***Competing interests***

490 The authors declare no competing financial interests and no conflict of interest.

491 ***Author contributions***

492 C. L.-S., J.-B. R. and D. A. C. conceived the experiment. C. L.-S., J.-B. R. and R. K.
493 performed the sample collection. C. L.-S. performed all experimental work. C. L.-S. and
494 C. C. performed the microbial and viral computational data analysis, respectively. C. L.-

495 S., C. C., J.-B. R. and D. A. C. participated in the interpretation of results and writing of
496 the manuscript. G. M.-K. provided logistical support and field advice in the Namib
497 Desert.

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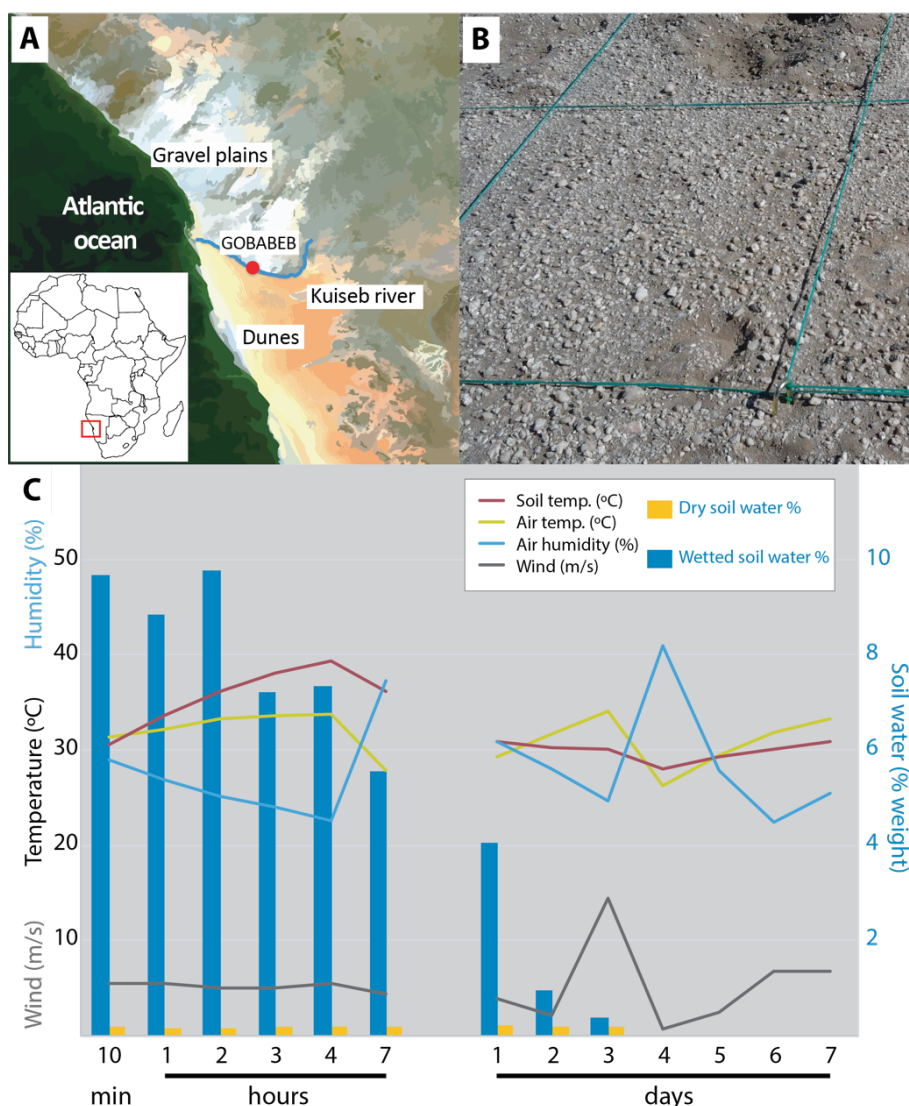
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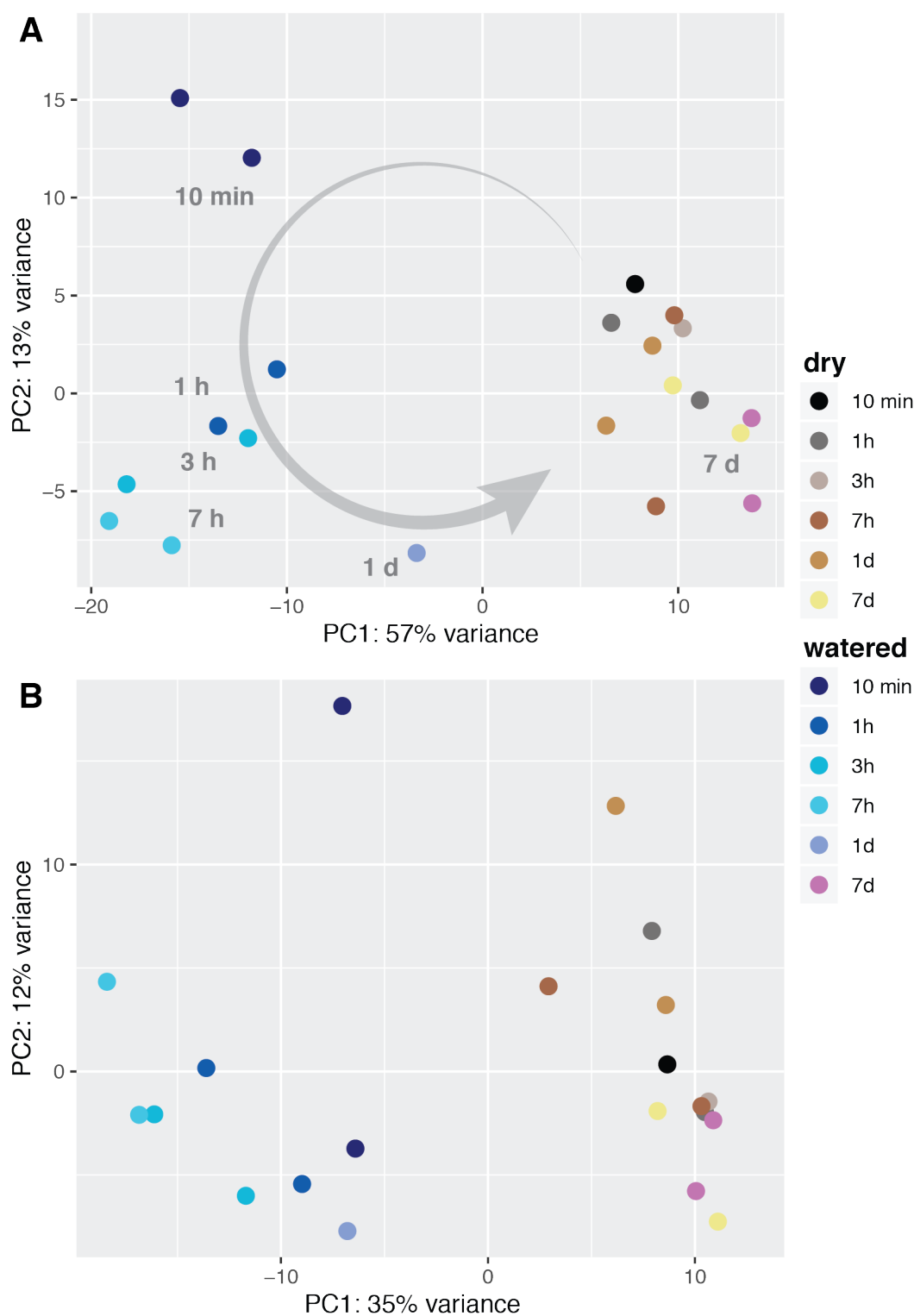
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821 **Figures**



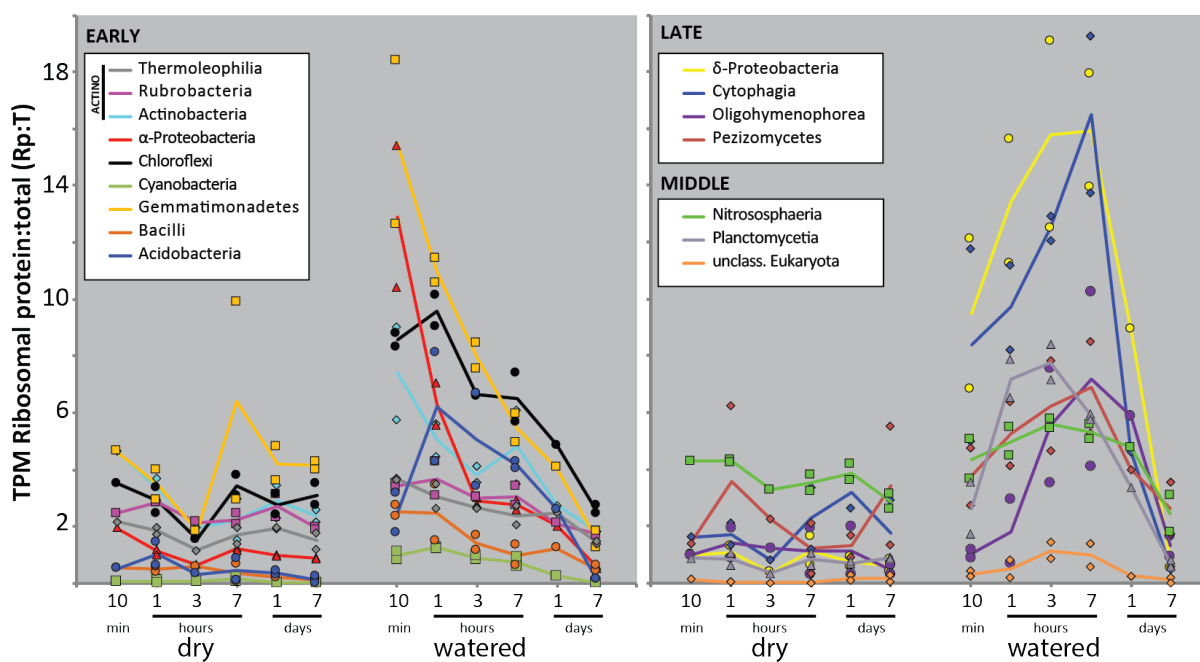
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823 **Figure 1:** Sample site and environmental conditions of the sampled soils. A) Map of
 824 the Namib Desert gravel plain location. Modified from European Space Agency,
 825 ESA/Envisat [CC BY-SA 3.0 IGO](https://creativecommons.org/licenses/by-sa/3.0/). B) View of a representative portion of the gravel plain
 826 sampling site. C) Environmental conditions over the experimental period: Air
 827 temperature, humidity and wind were recorded by the nearby Gobabeb meteorological
 828 station (Southern African Science Service Centre for Climate Change and Adaptive Land
 829 Management (SASSCAL), station 8893).



830

831 **Figure 2:** Principal components analysis of A) KEGG Orthologs (KO) annotated
832 transcripts, and B) taxonomic (family) assignment of transcripts. The arrow designates
833 the temporal transition pathway.



834

835 **Figure 3:** Ribosomal protein gene transcripts as a fraction of the total (Rp:T) among
 836 the most transcriptionally active microbial classes in the Namib soil community (>1%
 837 TPM average in dry or watered samples). Left panel shows *early* response classes whose
 838 Rp:T ratios peak within the first hour after watering. Right panel includes classes with
 839 Rp:T maxima 3 and 7 hours after watering (*middle* and *late* response taxa, respectively).

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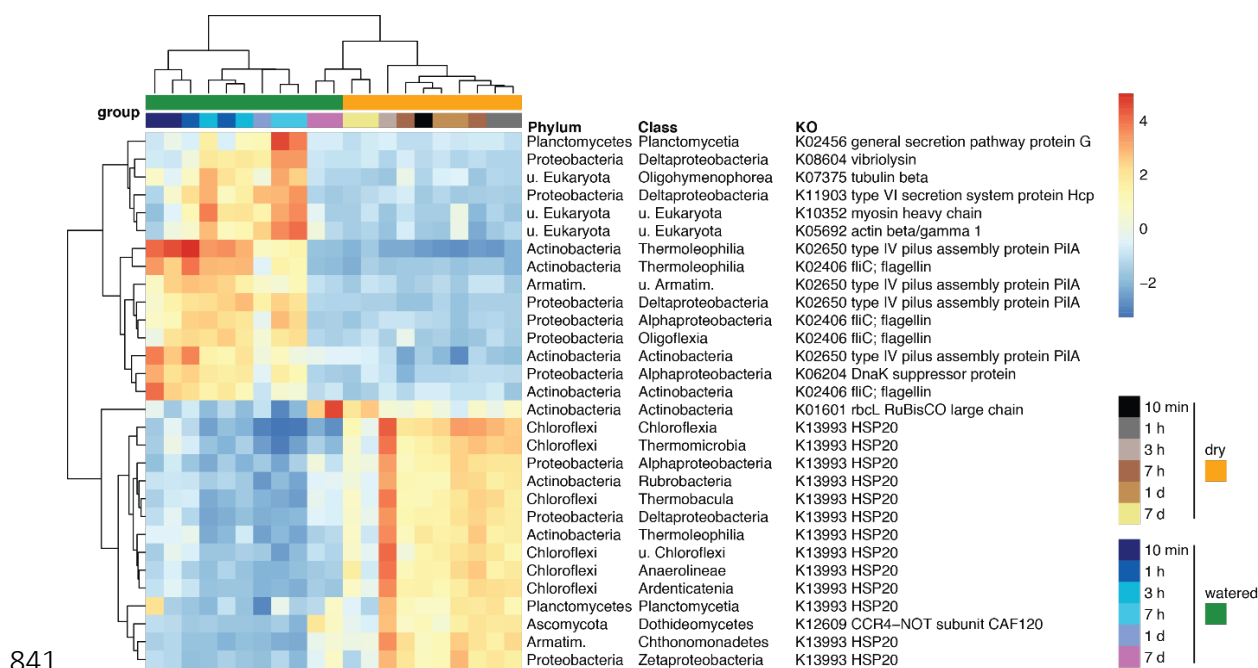


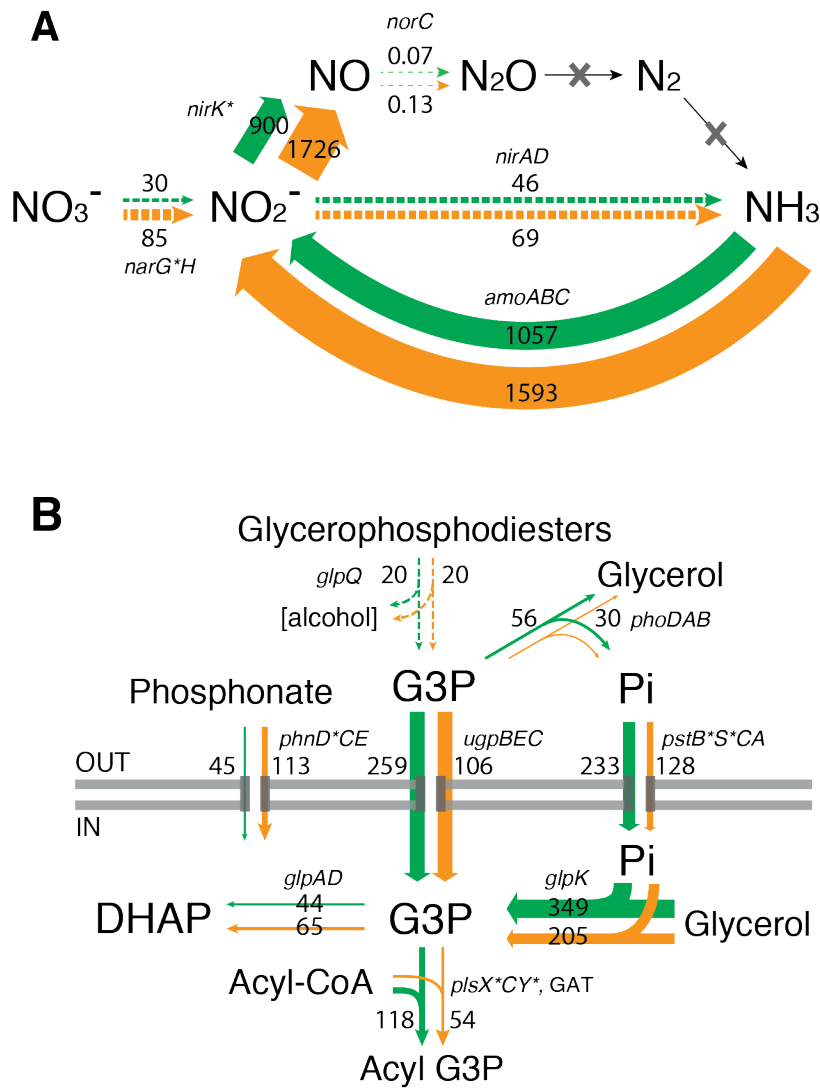
Figure 4: Temporal transcription patterns of the 30 most variable genes among

those with significantly differential transcription along the time series. Transcript data

was aggregated along taxonomic (class) as well as functional (KEGG Orthologs) groups

for the analysis. Values were normalized using the Variance Stabilizing Transformation

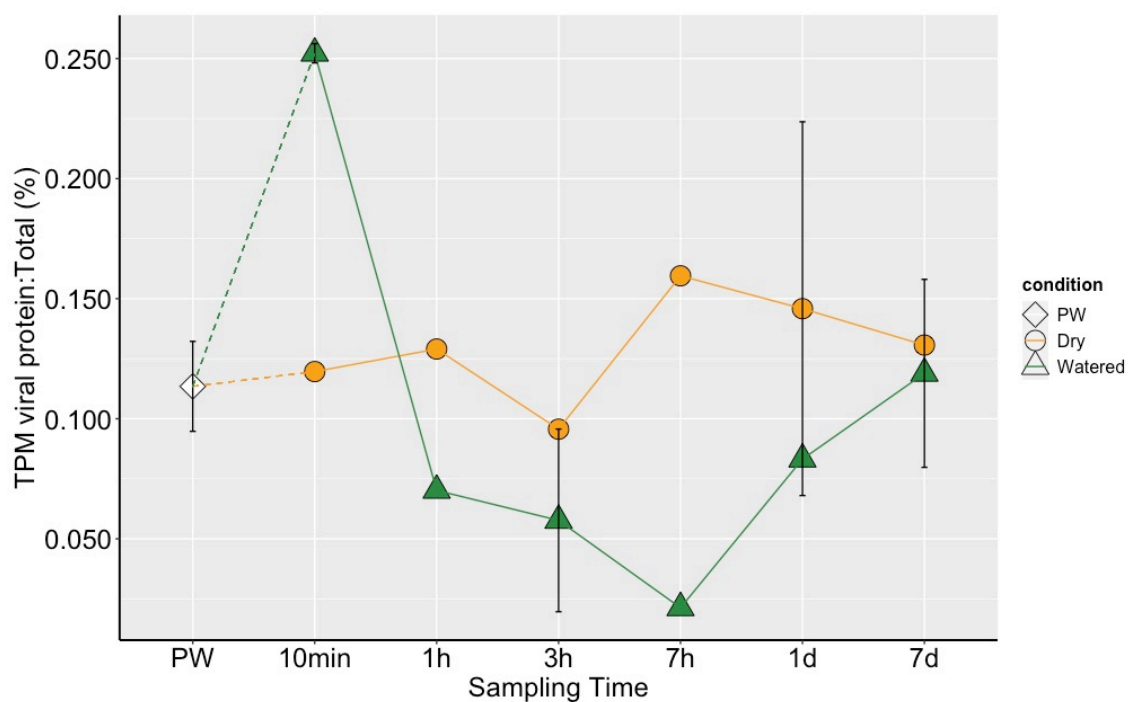
(*DESeq* R package).



847

848 **Figure 5:** A) Nitrogen cycling and B) phosphate assimilation system transcription in
 849 the surface soil microbial community. Average TPM measurements are provided for the
 850 dry plot samples (orange arrows) and wetted samples up to 7 hours post watering
 851 (green arrows). Asterisks indicate significant differentially transcribed KOs at community
 852 level at any point along the time series. Panel B figure modified from León-Sobriño *et*
 853 *al.*, 2019.

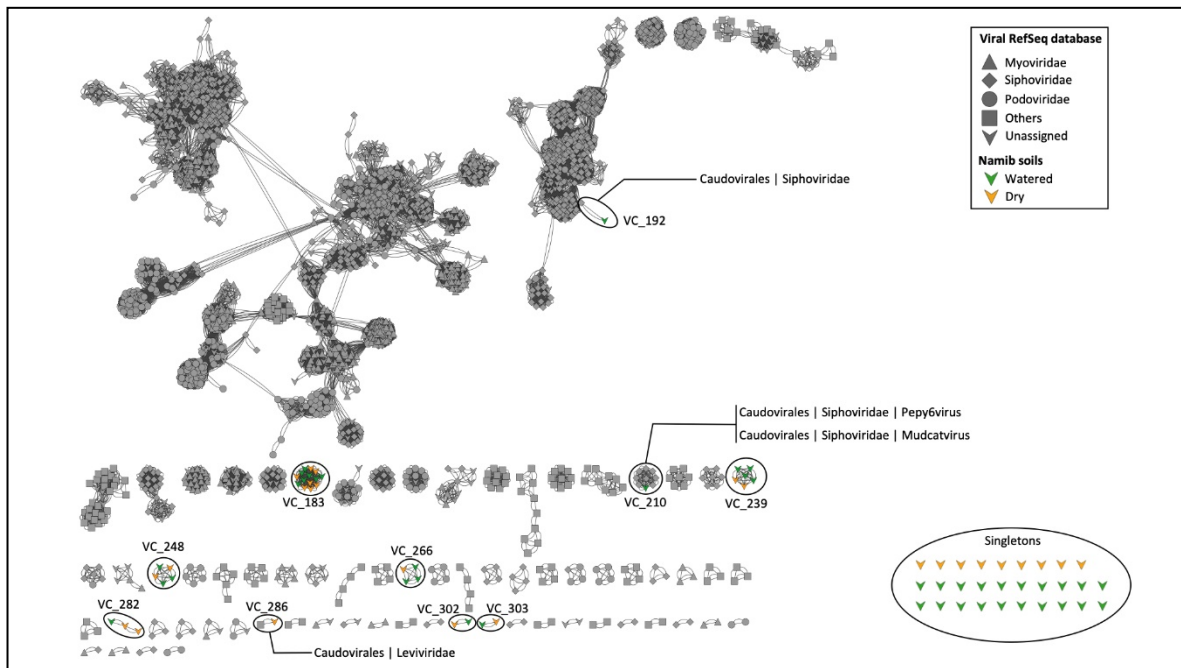
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856 **Figure 6.** Temporal changes in viral protein gene transcripts as a fraction of the total
857 (Vp:T ratios; %) for watered and dry (control) soil samples. The PW (Pre-watering) value
858 represents the mean \pm sd of Vp:T ratios in all control samples.

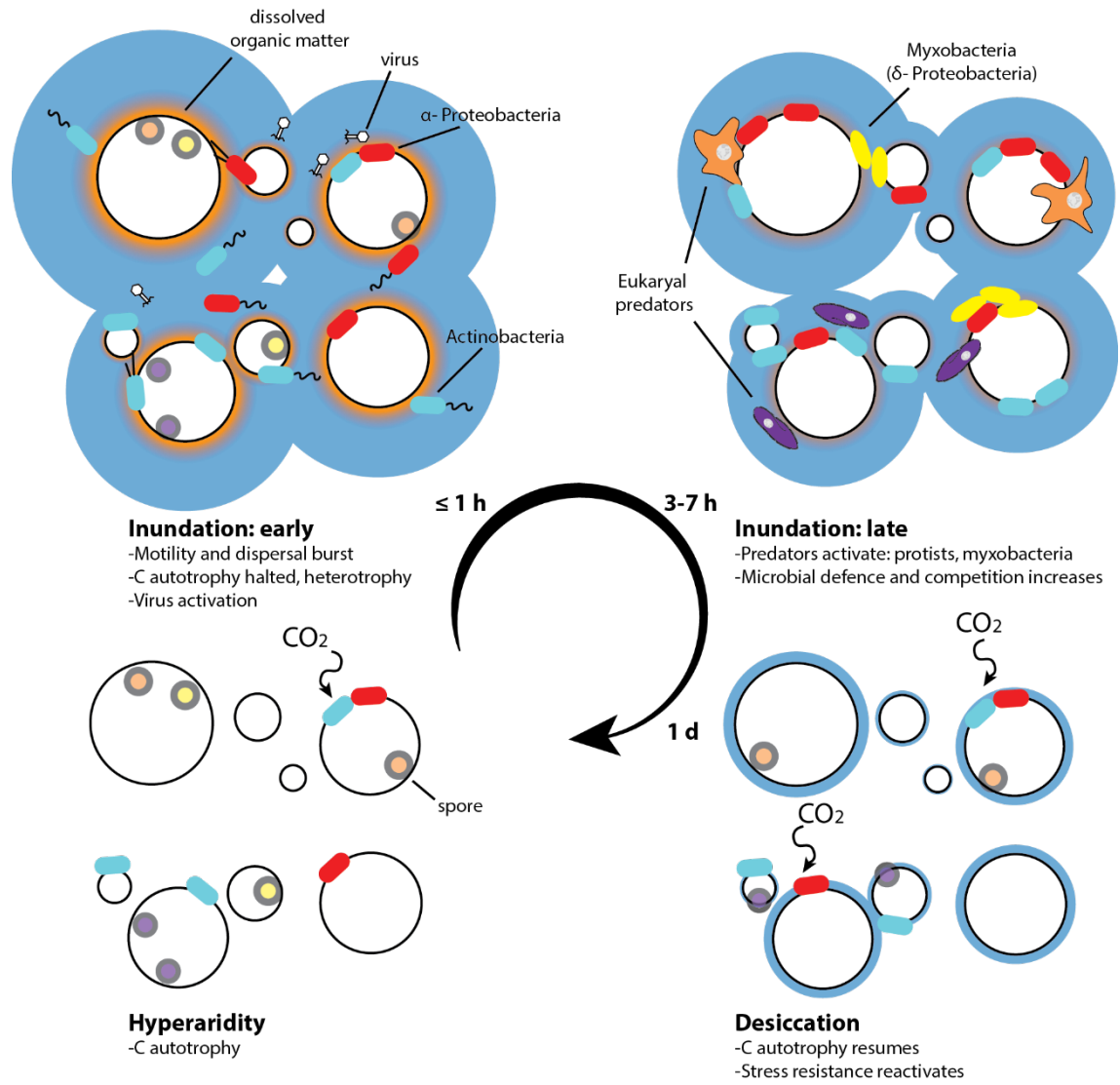
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860

861 **Figure 7.** Network analysis, relating Water Experiment (WE) phage sequences to
862 known viral sequences from the RefSeq Database. Circled clusters present viral contigs
863 identified in this study. Shapes indicate major viral families, RefSeq sequences are in
864 grey and WE contigs are colored (yellow and green for dry and watered samples,
865 respectively). Each node is depicted as a different shape, representing viruses belonging
866 to Myoviridae (triangle), Podoviridae (circle), Siphoviridae (diamond), or
867 uncharacterized viruses (V shape). Single shapes represent viral singletons identified in
868 this study. Edges (lines) between nodes indicate statistically weighted pairwise similarity
869 scores (see Materials and Methods) of ≥ 1 .

870



871

872 **Figure 8.** Response model of microbial communities to water events in hyperarid

873 desert soils.